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Journal of Experimental Marine Biology and Ecology 353 (2007) 58–67

Journal of EXPERIMENTAL MARINE BIOLOGY AND ECOLOGY

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Interspecific comparison of Cd bioaccumulation in European Pectinidae (*Chlamys varia* and *Pecten maximus*)

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Received 28 June 2007; received in revised form 30 August 2007; accepted 4 September 2007

Abstract

The uptake and loss kinetics of Cd were determined in two species of scallops from the European coasts, the variegated scallop *Chlamys varia* and the king scallop *Pecten maximus*, following exposures via seawater, phytoplankton and sediment using highly sensitive radiotracer techniques (^{109}Cd). Results indicate that, for seawater and dietary pathways, *C. varia* displays higher bioaccumulation capacities in terms of uptake rate from water and fraction absorbed from ingested food (assimilation efficiency) than *Pecten maximus*. Regarding sediment exposure, *P. maximus* displayed low steady-state Cd transfer factor ($TF_{SS} < 1$); however, once incorporated, a very large part of Cd transferred from sediment (92%) was strongly retained within *P. maximus* tissues.

Both species showed a high retention capacity for Cd (biological half-life, $T_{b1/2} > 4$ months), suggesting efficient mechanisms of detoxification and storage in both species. The digestive gland was found to be the main storage organ of Cd in the two scallops regardless of the exposure pathway. However, Cd was stored differently within this organ according to the species considered: 40% of the total Cd was found in the soluble cellular fraction in *C. varia* whereas this soluble fraction reached 80% for *P. maximus*. This suggests that the two species displayed different Cd detoxification/storage mechanisms.

Finally, the present study has determined the relative contribution of the different exposure pathways to global Cd bioaccumulation for the two scallop species. Results clearly show that for both species, food constitutes the major accumulation pathway, contributing for >99% and 84% of the global Cd bioaccumulation in *C. varia* and *P. maximus*, respectively. This work confirms the previous assumption, derived from a bibliographic overview, that dietary pathway plays a prevalent role in metal bioaccumulation in Pectinidae.

Keywords: Bivalves; Cadmium; Kinetics; Metal; Scallops; Subcellular Distribution

1. Introduction

Bivalves usually concentrate efficiently Cd from the surrounded environment (e.g. Eisler, 1985). Among them, Pectinidae can display very high concentrations of this non essential metal that is considered as one of the

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doi:10.1016/j.jembe.2007.09.001

most toxic ones. High levels of Cd in scallop tissues have been reported even for species from pristine and low-contaminated areas such as the Antarctic Ocean or the sub-polar Atlantic Ocean (Mauri et al., 1990; Viarengo et al., 1993; Bustamante and Miramand, 2004), suggesting that scallops have evolved a natural capacity to accumulate, detoxify and store this metal in their tissues. Investigations carried out in the field and in the laboratory have revealed the involvement of very efficient detoxification mechanisms. Indeed, the binding of Cd to high-affinity cytosolic proteins, lysosomes, and mineral concretions is well known to result in efficient Cd sequestration in Pectinidae (Carmichael and Fowler, 1981; Ballan-Dufrançais et al., 1985; Stone et al., 1986).

Even though field investigations have shown that Cd levels are influenced by various factors such as geographical origin, season, size and sexual maturity (Bryan, 1973; Evtushenko et al., 1990; Mauri et al., 1990; Bustamante and Miramand, 2004, 2005a), very little is known on the dynamics of Cd bioaccumulation and retention in this family. To the best of our knowledge, no study has described the Cd accumulation in Pectinidae exposed via different pathways and its depuration using environmentally realistic metal levels. For example the earlier study by Eisler et al. (1972) exposed *Aquipecten irradians* to 10 ppm Cd, a concentration with toxic consequences (Gould et al., 1988) and therefore unlikely to produce a typical accumulation pattern for Cd. In natural conditions, scallops are exposed to metal through seawater and food pathways, sediment potentially contributing to either or both. It is therefore necessary to investigate separately these different exposure pathways to understand their relative contribution in the global accumulation of the metal (Fowler, 1982).

Seawater has been often considered as the main source of metal intake for marine organisms (e.g., Janssen and Scholz, 1979; Borchardt, 1983; Riisgard et al., 1987); however the role of the particulate phase, mainly food, is now recognized to be of primary importance for a large range of taxa (e.g., Warnau et al., 1996, 1999; Reinfelder et al., 1998; Wang and Fisher, 1999). In the case of Pectinidae, it has been suggested that food could be the major route of Cd intake on the basis of elevated metal concentrations found in the digestive gland (Palmer and Rand, 1977; Uthe and Chou, 1987; Bustamante and Miramand, 2005a). However, it appears necessary to confirm this assumption as the contribution of the dissolved phase could also lead to high metal concentrations in the detoxification and storage organs (e.g., Borchardt, 1983).

Therefore, the present work investigated uptake and loss kinetics of Cd in two species of scallops, *Chlamys varia* and *Pecten maximus* exposed through seawater, food and/or sediment, depending of their different living habitats- only seawater and food for *C. varia* and all pathways for *P. maximus* which is living buried in the bottom sediment and is able to ingest large particles (Mikulich and Tsikhon-Lukamina, 1981; Shumway et al., 1987). The use of highly sensitive radiotracer techniques allowed studying bioaccumulation mechanisms at realistic Cd levels encountered in the field. Three levels of biological organization were considered in this study, the whole individual, the different organs and the subcellular fractions of the digestive gland cells, in order to evaluate the biokinetic parameters of the accumulation, the distribution among the body compartments and the cellular forms of storage in the digestive gland, respectively. Finally, we used a bioaccumulation model to determine the relative contribution of the different exposure pathways of Cd for both species.

2. Materials and methods

2.1. Sampling

In spring 2004 and 2005, one hundred variegated scallops *C. varia* and seventy king scallops *P. maximus* were collected on the Atlantic coast (Pertuis Breton, Charente-Maritime) by SCUBA diving. They were carefully transported to IAEA-MEL premises in Monaco and were acclimatized to laboratory conditions for 4 weeks (constantly aerated open circuit aquarium; flux: 50 l h^{-1}; salinity: 36 p.s.u.; temperature: 17±0.5 °C; pH: 8.0±0.1; light/dark cycle: 12 h/12 h) prior to experimentations. During this period, scallops were fed daily an algal mixed diet (*Isochrysis galbana, Skeletonema costatum*).

2.2. Radiotracer and counting

Uptake and loss kinetics of ^{109}Cd in scallop species were determined using a high specific activity radiotracer purchased from Isotope Product Lab (^{109}Cd as $CdCl_2$ in 0.1 M HCl, $T_{1/2}$=426.6 d). The tracer was counted using a high-resolution γ-spectrometer system composed of four Germanium (N- or P-type) detectors (EGNC 33-195-R, Canberra® and Eurysis®) connected to a multichannel analyser (Intergamma, Intertechnique). The radioactivity was determined by comparison with standards of known activity and of appropriate geometry. Measurements were corrected for counting efficiency and physical radioactive decay. The counting time was adjusted to obtain a propagated counting error less than 5%.

2.3. Seawater exposure

Twenty three *C. varia* and 23 *P. maximus (*average weight±SD: 30±7 g and 208±46 g, respectively) were placed in a 70-l glass aquarium (constantly aerated closed circuit aquarium; salinity: 36 p.s.u.; temperature: 17±0.5 °C; pH: 8.0±0.1; light/dark cycle: 12 h/12 h) and exposed for 7 d to ^{109}Cd dissolved in seawater (2 kBq l^{-1}). No change in pH was detectable after the tracer addition. Spiked seawater was renewed twice a day the first two days and then daily in order to keep radioactivity in seawater constant. Activity of the ^{109}Cd in seawater was checked before and after each spike renewal, yielding time-integrated activities of 2.1± 0.2 kBq l^{-1}.

Nine scallops of each species were collected at different time intervals and were whole-body radioanalyzed alive (same identified individual each time). At the end of the 7-day exposure period, 5 scallops of each species were sacrificed and dissected. Shell, digestive gland, kidneys, gills, gonad, mantle, intestine, adductor muscle and the remaining soft tissues were separated and radioanalyzed in order to assess the ^{109}Cd body distribution. The remaining scallops were then placed in non contaminating conditions (constantly aerated open circuit; flux: 50 l h^{-1}; salinity: 36 p.s.u.; temperature: 17± 0.5 °C; pH: 8.0±0.1; light/dark cycle: 12 h/12 h) for 36 d and nine individuals of each species were regularly radioanalyzed alive in order to follow the loss of ^{109}Cd from the scallops. Four scallops were collected at the end of the depuration period and dissected into several body compartments as previously described.

2.4. Food exposure

The Haptophyceae *Isochrisis galbana* was used to study ^{109}Cd transfer to scallops through their diet. Phytoplankton cells were exposed to 4.8 kBq l^{-1} ^{109}Cd during their growing phase (7 d). After that period, phytoplankton medium was filtrated (1 μm-mesh size; Osmonic filters), and then resuspended in a 70-l aquarium (constantly aerated closed-circuit; salinity: 36 p.s.u.; temperature: 17±0.5 °C; pH: 8.0±0.1; light/dark cycle: 12 h/12 h) where six *C. varia* and six *P. maximus* (average weight±SD: 17±5 g and 127±14 g, respectively) were placed for one week before the feeding experiment. The radioactivity of the labelled *I. galbana* was γ-counted before and after the filtration. Scallops were allowed to feed on radiolabelled *I. galbana* for 2 h (cell concentration – $5 \cdot 10^4$ cell ml^{-1} – was selected to avoid pseudofeces production). After the feeding period, all scallops were γ-counted and flowing seawater conditions (50 l h^{-1}) were restored in the aquarium. Individuals were then whole-body γ-counted alive at different time intervals to follow the loss kinetics of ^{109}Cd. Four individuals were collected after 16 (*P. maximu*s) and 30 d (*C. varia*) of depuration, and dissected to determine the ^{109}Cd tissue distribution among the different body compartments (shell, digestive gland, kidneys, gills, gonad, mantle, intestine, adductor muscle and the rest of soft tissues) and among the subcellular fraction of the digestive gland (see below).

2.5. Sediment exposure

Since *P. maximus* is living buried into the sediment whereas *C. varia* is fixed on rocks, Cd exposure through sediment was only assayed for *P. maximus*. Sediment was collected in Wimereux (North-Atlantic coast of France). Sediment grain size distribution was measured on a Mastersizer micro and the evaluation of the dry/wet weight ratio was calculated after freeze drying in a LABCONCO Freezone18. Aerated sediment (9 kg) was placed in plastic bottle, exposed to ^{109}Cd (516 kBq) for 6 d with constant agitation, then used to form a homogeneous sediment layer of 4 cm height in a 20-l aquarium. Weakly bound ^{109}Cd was allowed to leach overnight under flowing seawater (50 l h^{-1}). Ten *P. maximus* (average weight±SD: 118±5 g) were then placed for 13 d in the aquarium (constantly aerated open circuit; flux: 50 l h^{-1}; salinity: 36 p.s.u.; temperature: 17±0.5 °C; pH: 8.0±0.1; light/dark cycle: 12 h/12 h). Six individuals as well as sediment aliquots were regularly radioanalyzed during the experiment duration. Activity of ^{109}Cd in sediment was constant all along the exposure period (24.2± 1.9 Bq g^{-1} wet wt). At the end of the uptake period, 4 scallops were collected, dissected (shell, digestive gland, kidneys, gills, gonad, mantle, intestine, adductor muscle and the rest of soft tissues), weighed and γ-counted in order to determine the radiotracer distribution among the body compartments. The remaining individuals were transferred for 49 d to a new 20-l aquarium containing non contaminated sediment with flowing seawater and they were regularly radioanalyzed to follow ^{109}Cd loss kinetics. Also, ^{109}Cd activity in sediment was regularly measured in order to ascertain that no contamination of the clean sediment occurred through ^{109}Cd recycling (for security, the whole sediment layer was renewed anyway after one week). At the end of the loss period, 4 scallops were collected and dissected as described above to determine ^{109}Cd body distribution and its subcellular distribution in the digestive gland.

2.6. Subcellular distribution

For all the experiments, the digestive gland of both scallop species were considered to assess the partitioning of ^{109}Cd between soluble and insoluble fractions as described by Bustamante & Miramand (2005b). Briefly, part of digestive gland were homogenized individually with a mortar and pestle on ice with 10 ml of 0.02 M Tris–HCl buffer, 0.25 M sucrose, 1 mM phenylmethylsulfonylfluoride (PMSF, as protease inhibitor), at pH 8.6. The homogenates were centrifuged at 80,000 *g* for 1 h at 5 °C in a Sorvall RC28S ultracentrifuge to separate particle-free supernatant (cytosol; soluble fraction) from the pellet (insoluble fraction). Homogenate aliquots, cytosols, and pellets were then radioanalyzed.

2.7. Data analysis

Uptake of the radioisotope was expressed in term of concentration factors (CF: ratio between the ^{109}Cd activity in scallops – Bq g^{-1} wet wt – and time-integrated activity in the seawater — Bq g^{-1}) over time for the seawater exposure and in term of transfer factors (TF: ratio between the ^{109}Cd activity in scallops – Bq g^{-1} wet wt – and time-integrated activity in the sediment — Bq g^{-1}) over time for the sediment exposure of *P. maximus*. Uptake kinetics of ^{109}Cd in whole-body scallops were fitted using a simple exponential kinetic model (Eq. (1)) for the sediment exposure (Statistica® 6) and using a linear model for the seawater exposure (Eq. (2)):

$$CF_t = CF_{ss}\left(1 - e^{-k_e t}\right) \quad (1)$$

$$CF_t = k_u t \quad (2)$$

where CF_t and CF_{ss} ($CF_{ss} = k_u/k_e$) are the concentration factors at time t (d) and at steady state, respectively; k_u and k_e are the uptake and loss rate constants (d^{-1}), respectively (Whicker and Schultz, 1982).

Depuration of Cd (seawater, food and sediment experiments) was expressed in terms of percentage of remaining radioactivity (radioactivity at time t divided by initial radioactivity measured in scallops at the beginning of the decontamination period * 100). The percentages of remaining activity were plotted against time and loss kinetics were described by a double-component exponential model (Eq. (3)):

$$A_t = A_{0s}e^{-k_{es}t} + A_{0l}e^{-k_{el}t} \quad (3)$$

where A_t and A_0 are the remaining activities (%) at time t (d) and 0, respectively; k_e is the depuration rate constant (d^{-1}); 's' and 'l' are the subscripts for the

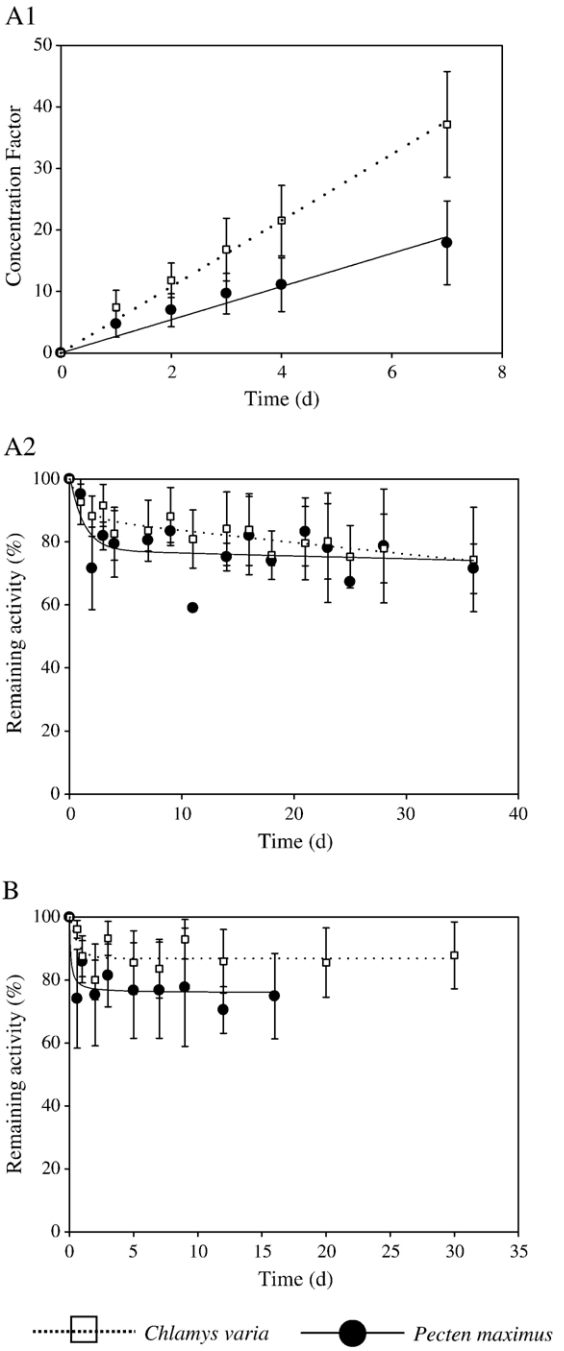

Fig. 1. *Chlamys varia* and *Pecten maximus.* Uptake and loss kinetics of ^{109}Cd in scallops exposed for 7 d via seawater (uptake kinetics A1; Concentration Factors; mean±SD; $n=9$), then maintained for 36 d in non contaminated conditions (loss kinetics A2; Remaining activity (%); mean±SD; $n=9$) and after a 2-h feeding on radiolabelled phytoplankton *Isochrysis galbana* (loss kinetics B; Remaining activity (%); mean±SD; $n=6$ *C. varia* and $n=9$ *P. maximus*).

'short-lived' and 'long-lived' components. For each exponential component (s and l), a biological half-life can be calculated ($T_{b1/2s}$ and $T_{b1/2l}$) from the corresponding depuration rate constant (k_{es} and k_{el}, respectively) according to the relation $T_{b1/2}=\ln 2/k_e$ (Warnau et al., 1996). Regarding feeding experiments, the 'long-lived' exponential term describes the fraction of the radiotracer ingested with food that is actually absorbed by the organism (Warnau et al., 1996). The corresponding A_{0l} represents the assimilation efficiency (AE) of the considered radiotracer. The best fitting regression models were selected according to highest determination coefficient and examination of residuals. The level of significance for statistical analysis was always set at $\alpha<0.05$.

3. Results

3.1. Seawater exposure

Uptake of ^{109}Cd in whole-body *C. varia* and *P. maximus* displayed linear kinetics ($r^2=0.85$ and 0.66, respectively; see Fig. 1). The values estimated for the kinetic parameters and their associated statistics are presented in Table 1. The concentration factors measured at the end of the uptake period (CF_{7d}) of ^{109}Cd were 37±9 in *C. varia* and 18±7 in *P. maximus* (Table 2). Calculated CF_{7d} for the different organs indicated that ^{109}Cd was concentrated selectively in each species, according to the following order:

C. varia: kidneys (928 ± 547) > digestive gland (322 ± 175) ≈ gills (277 ± 102) ≈ foot (265 ± 74) ≈ rest of soft tissues (258 ± 56) > gonad, mantle, intestine and adductor muscle (≤ 53 ± 11);

P. maximus: kidneys (690 ± 402) ≈ digestive gland (659 ± 227) > gills (175 ± 13) > other tissues (≤ 78 ± 33).

In terms of body distribution, ^{109}Cd was mainly found in the digestive gland and in the gills (~30 and 20% of total body load, respectively) for both species. At the end of the uptake experiment, the ^{109}Cd tissue distribution shows a similar pattern ($p_{G\text{-test}}>0.40$) between *C. varia* and *P. maximus*, with the digestive gland and gills accounting for more than 60% of the total Cd load (Table 2).

After the exposure period, non-contaminating conditions were restored and loss kinetics of ^{109}Cd were followed for 36 d. The whole-body loss kinetics of ^{109}Cd in *C. varia* and *P. maximus* were best described by a two-component exponential model (Fig. 1 and Table 1). The major part of ^{109}Cd was efficiently absorbed in *C. varia* and *P. maximus* ($A_{0l}>77\%$). The estimated loss rate constant of the long-lived components (k_{el}) for *C. varia* was low, i.e. 0.005±0.001 and, consequently, the derived biological half-life reached 145±45 d (Table 1). In the case of *P. maximus*, the loss rate constant was not significantly different from 0 ($p>0.05$), and the related $T_{b1/2l}$ of ^{109}Cd may thus be considered as infinite.

After 36 d of depuration, the body distribution of ^{109}Cd displayed a similar pattern than the one observed at the end of the exposure period (Table 2). However, it is striking to note that the ^{109}Cd activity in the digestive gland of *C. varia* and *P. maximus* remained relatively constant throughout the depuration duration within the two species, i.e. from 680±369 Bq g^{-1} to 549±255 Bq g^{-1} for *C. varia* and from 1,392±479 Bq g^{-1} to 1,491±316 Bq g^{-1} for *P. maximus*, suggesting either a lack of Cd loss from the digestive gland during this period or a redistribution of the radioisotope from

Table 1
Chlamys varia and *Pecten maximus*

Experiment	Species	Uptake			Loss				
		CF_{ss}/TF_{ss}±ASE	k_u±ASE	r^2	A_{0s}±ASE	$T_{b1/2s}$±ASE	A_{0l}±ASE	$T_{b1/2l}$±ASE	r^2
1) Seawater	*C. varia*	–	5.4±0.2[d]	0.85	12.2±3.8[b]	0.8	87.8±2.4[d]	145±45[b]	0.31
	P. maximus	–	2.7±0.1[d]	0.66	23.4±5.7[c]	1.1	77.1±4.8[d]	913	0.49
2) Feeding	*C. varia*	–	–	–	14.5±4.1[c]	0.4	85.8±2.1	989	0.21
	P. maximus	–	–	–	20.5±6.1[b]	0.02	79.5±3.7[d]	138	0.37
3) Sediment	*P. maximus*	0.034±0.002[d]	0.014±0.002[d]	0.62	NC	NC	92[d]	NC	

Whole-body uptake and loss kinetic parameters of ^{109}Cd following different exposure experiments:
1) 7-d exposure via seawater ($n=9$) followed by 36 d of depuration ($n=9$);
2) 2-h feeding on radiolabelled *Isochrysis galbana* followed by a depuration period of 16 d (*P. maximus*, $n=6$) or 30 d (*C. varia*, $n=6$);
3) 13-d exposure of *P. maximus* via the sediments ($n=8$) followed by 31 d of depuration ($n=8$).
Uptake parameters: CF_{ss}/TF_{ss} concentration and transfer factors at steady state; k_u: uptake rate constant (d^{-1}).
Depuration parameters: A_{0s} and A_{0l}: activity (%) lost according to the short-and the long-lived exponential component, respectively; $T_{b1/2}$: biological half-life (d). ASE: asymptotic standard error; r^2: determination coefficient of the uptake or loss kinetics.
Probability -*p*- of the model adjustment: [a]$p<0.05$, [b]$p<0.01$, [c]$p<0.001$, [d]$p<0.0001$; NC: not calculated.

Table 2
Chlamys varia and *Pecten maximus*

Species compartments	Seawater contamination			Food contamination
	Uptake (7 d, $n=5$)		Loss (36 d, $n=4$)	Loss ($n=5$)
	Concentration factor	Distribution (%)	Distribution (%)	Distribution (%)
Chlamys varia				
Digestive gland	322±175	33±14	41±18	97±1
Gills	277±102	30±9	23±6	<1
Kidneys	928±547	13±6	15±8	<1
Intestine	23±7	<1	1±1	<1
Gonad	45±65	1±1	1±1	1±0
Foot	265±74	3±1	2±0	<1
Mantle	53±11	12±4	10±6	<1
Adductor muscle	21±6	4±1	5±3	<1
Remaining tissues	258±56	5±1	2±0	0±1
Whole body	37±9			
Pecten maximus				
Digestive gland	659±227	38±10	49±5	82±19
Gills	175±13	28±11	19±2	1±0
Kidneys	690±402	10±4	12±4	6±12
Intestine	16±3	<1	<1	1±1
Gonad	18±10	2±2	2±1	9±17
Foot	13±5	<1	<1	1±1
Mantle	28±5	11±2	10±7	<1
Adductor muscle	18±7	9±3	7±1	<1
Remaining tissues	78±33	2±0	1±0	1±0
Whole body	18±7			

Concentration Factors (mean CF±SD) and tissue distribution (mean %±SD) of ^{109}Cd during seawater (end of exposure and depuration periods) and feeding experiments (16 and 30 d after feeding for *P. maximus* and *C. varia*, respectively).

the tissues in contact with seawater towards this storage organ.

3.2. Dietary exposure

The loss kinetics of ^{109}Cd ingested with food in both *C. varia* and *P. maximus* were best fitted using a double exponential model (Fig. 1 and Table 1). *C. varia* displayed a higher assimilation efficiency (AE>86%) than *P. maximus* (AE>80%). However, in both species, the depuration rate constant, k_{el}, were not significantly different from 0 ($p>0.39$), and therefore the derived $T_{b1/2l}$ were infinite.

At the end of the depuration period, the digestive gland contained the main part of ^{109}Cd, i.e. 97% for *C. varia* and 82% for *P. maximus* (Table 2).

3.3. Sediment exposure

Sediment used in the experiment was mainly (95.8%) composed of grains which size ranged from 76 to 302 μm and its dry/wet wt ratio was 0.80.

Whole-body uptake kinetics of sediment-bound ^{109}Cd in *P. maximus* was best fitted by a single exponential model (Table 1). TF reached steady-state equilibrium within the 2 weeks of exposure (estimated $TF_{ss}=0.034\pm0.002$). Among the different body compartments, the highest TF_{13d} was found in the digestive gland (3.35±1.68; Table 3). This organ also contained the main

Table 3
Pecten maximus

Compartments	Uptake phase		Loss phase
	Transfer factor	Distribution (%)	Distribution (%)
Digestive gland	3.35±1.68	78±10	80±10
Gills	0.05±0.04	4±3	6±1
Kidneys	0.12±0.04	1±1	<1
Intestine	0.09±0.05	<1	<1
Gonad	0.06±0.05	1±0	<1
Foot	0.03±0.01	<1	<1
Mantle	0.06±0.02	14±8	12±10
Adductor muscle	0.00±0.00	1±1	<1
Remaining tissues	0.06±0.05	1±1	<1
Whole body	0.04±0.01		

Transfer factors (mean TF±SD; $n=4$) of ^{109}Cd after a 13-d exposure via sediment and tissue distribution (mean %±SD) of ^{109}Cd at the end of the 13-d exposure and 31-d depuration period ($n=5$).

fraction of the total ^{109}Cd body burden (i.e. 78%; Table 3). The body compartment containing the second highest proportion was the mantle (14% of total ^{109}Cd body burden).

The ^{109}Cd whole-body loss kinetics could not be described accurately by the exponential models; therefore a linear regression ($Y=aX+b$) was applied in order to estimate the radiotracer retention. The results showed that 92% of the accumulated ^{109}Cd were efficiently incorporated in *P. maximus* tissues, with a biological half-life not significantly different from infinite (Table 1). At the end of the depuration period (31 d) the body distribution of ^{109}Cd was identical to that at the end of the exposure period (Table 3), with the highest proportion of ^{109}Cd located in the digestive gland (≈80%), followed by the mantle (≈12–14%). In addition, the ^{109}Cd activities were similar in the two latter tissues at the end of exposure and depuration periods, viz. 81±41 and 85±18 Bq g^{-1} in the digestive gland and 1.4±0.4 and 1.5±1.4 Bq g^{-1} in the mantle.

3.4. Subcellular distribution

Examination of subcellular distributions indicated that, whatever the contamination pathway (i.e., seawater, food or sediment) and the sampling period (i.e., end of uptake or end of loss period), *P. maximus* stored the major part of the cellular ^{109}Cd in the soluble fraction (from 70 to 85%). In contrast, the radiotracer was mainly bound to insoluble compounds in *C. varia* (Fig. 2).

4. Discussion

Pectinidae are an important marine resource which are both fished and cultured for human consumption (Ansell et al., 1991; Waller 1991). Hence, the intake of contaminants such as metals by Man through scallop consumption is a matter of concern. Indeed, Pectinidae are well known for their capacity of accumulating high levels of metals, and especially Cd, in their tissues (Brooks and Rumsby, 1965; Bryan, 1973; Bustamante and Miramand, 2004, 2005b). Interestingly, this high bioaccumulation potential for Cd is not specific to anthropogenic contamination since scallops from the Antarctic Ocean have high Cd levels compare to temperate species living in the coastal waters of industrialised countries (Mauri et al., 1990; Viarengo et al., 1993).

Several field studies assumed that food would be the main intake pathway of Cd in scallops as high metal levels are always found in the digestive gland (Palmer and Rand, 1977; Uthe and Chou, 1987; Bustamante and Miramand, 2005a). However, the contribution of the dissolved phase is difficult to ascertain in the field as this route can lead to a significant uptake of Cd and to its redistribution towards storage tissues such as the digestive gland. Therefore, there is a need to assess the relative importance of dissolved and particulate Cd pathways in order to better understand their respective contributions, as well as to evaluate the retention mechanisms leading to the high Cd levels measured in scallop tissues.

The experimental exposure of *C. varia* and *P. maximus* to ^{109}Cd via seawater confirmed their ability

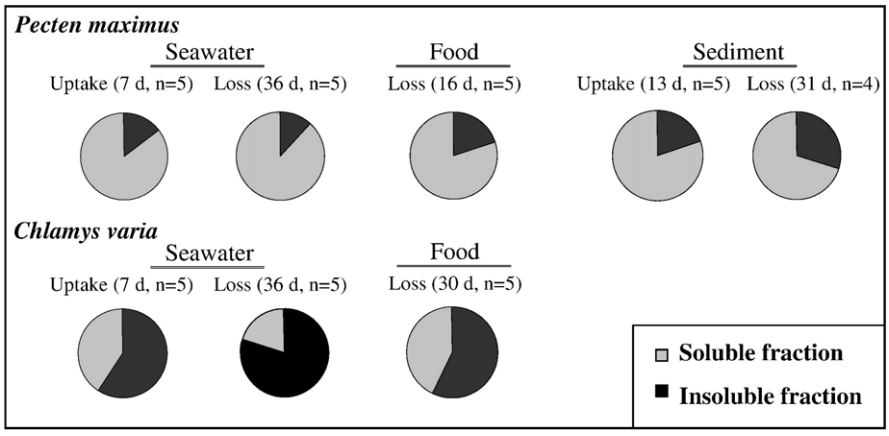

Fig. 2. *Chlamys varia* and *Pecten maximus.* Subcellular distribution of ^{109}Cd in the digestive gland cells following different exposure experiments: (1) 7-d exposure via seawater followed by 36 d of depuration; (2) 2-h feeding on radiolabelled *Isochrysis galbana* followed by a depuration period of 16 d (*P. maximus*) or 30 d (*C. varia*); (3) 13-d exposure of *P. maximus* via the sediments followed by 31 d of depuration.

to concentrate Cd from the dissolved phase, as previously shown using elevated exposure levels of stable Cd (Eisler et al., 1972; Carmichael and Fowler, 1981). Indeed, after only 7 days of exposure to the dissolved radiotracer, both scallop species exhibited high whole-body concentration factors (CFs), with 37 ± 9 for *C. varia* and 18 ± 7 for *P. maximus*. This difference in CF between the two species exposed to the same contamination conditions is related (1) to a higher Cd uptake rate (uptake rate constant: 5.4 vs 2.7) and (2) secondarily, to a higher assimilated fraction (87.8 vs 77.1) in *C. varia* compared to *P. maximus* (Table 1). However in the specimens collected from the field, *C. varia* displayed typically lower Cd concentrations than *P. maximus* (Palmer and Rand, 1977; Uthe and Chou, 1987; Bustamante and Miramand, 2005a). This would suggest that *C. varia* has far more limited capacities of Cd storage than *P. maximus*.

Considering the tissues separately, the organs involved in respiration (i.e. gills), excretion (i.e. kidneys) and digestion (i.e. digestive gland) displayed higher CFs compared to other body compartments in *P. maximus,* whereas the foot and the compartment "remaining soft tissues" also showed elevated CFs in *C. varia* (see Table 2). However, in terms of distribution among tissues and organs, Cd was mainly located in the digestive gland, the gills, the kidney and the mantle in both species, the digestive gland containing more than 30% of the whole body burden of ^{109}Cd (Table 2). These results strongly suggest the occurrence of efficient redistribution mechanisms towards the tissues involved in the detoxification, storage and excretion processes, i.e. the kidneys and the digestive gland (e.g., Carmichael and Fowler, 1981; Ballan-Dufrançais et al., 1985; Stone et al., 1986). It is also striking to note the difference between both species concerning the Cd CF in the foot that reached elevated values in *C. varia* (Table 2). In this species, the foot is well developed and contains a byssal gland which main role is to produce the byssus to stick to rocky substrates whereas *P. maximus* does not produce byssus as it lives buried in the sediment. Byssus is known to play a role in the elimination of metals from bivalves (Szefer et al., 2006), it is therefore likely that some metals are transferred from the soft tissues and concentrated in the byssus rather than merely adsorbed onto its surface from seawater. However, in the case of Cd, previous studies on mussels suggested that this metal is derived mainly from seawater (Coombs and Keller, 1981; Nicholson and Szefer, 2003). The present study was not designed to address this specific issue and our results do allow supporting internal transfer or waterborne origin of Cd in the byssus. However, further specifically-designed studies using sensitive radiotracer techniques could bring most interesting information on the origin of byssal Cd.

It is noteworthy that the Cd distribution pattern among the tissues was similar after 7 d of seawater exposure and after 36 d of depuration for both species (Table 2). Similarly, the subcellular distribution of Cd was identical at both times for *P. maximus*, with more than 80% in the soluble fraction of the digestive gland cells (Fig. 2). Taking into account the relatively long biological half-life of Cd in *P. maximus*, this result indicates that the metal is mainly bound to soluble compounds involved in the storage of this metal. The implication of metallothionein-like proteins in Cd detoxification and storage in the digestive gland is well documented in Pectinidae (e.g., Stone et al., 1986; Evtushenko et al., 1990; Bustamante and Miramand, 2005b). However, in *C. varia*, Cd was mainly bound to insoluble compounds (from 59 to 80%; see Fig. 2), suggesting a time-limited role of the soluble metallo-proteins when the metal enters through the dissolved route (as well as via the food as similar results were found for the dietary exposure; see Fig. 2). Such a predominant interaction of Cd with the insoluble cellular fraction in the digestive gland is not a common observation among Pectinidae but has already been shown in some species (e.g., *Adamussium colbecki*; Viarengo et al., 1993) and would be due to the fact that, among insoluble cellular components (i.e., organelles, membranes and granules), the lysosomal system can play a major role in Cd detoxification (by trapping) and excretion (Ballan-Dufrançais et al., 1985; Marigómez et al., 2002).

After exposure to sediment-bound Cd, *P. maximus* exhibited very low transfer factors (viz., $TF_{ss} = 0.034 \pm 0.009$), indicating that direct contamination due to burying into sediment would represent a minor Cd uptake pathway in this species. However, at the end of the exposure period, 80% of the incorporated metal was found in the digestive gland, which displayed a TF higher than 3 (Table 3). As this organ is not in direct contact with the sediment, it is suggested that either (1) the radiotracer was progressively translocated from the tissues in direct contact with sediment and pore water to the digestive gland and/or (2) *P. maximus* was able to ingest sediment grains. Although sediment grains were never observed in the valves or in the digestive system in the many dissections carried out during this study, this latter hypothesis would be plausible as scallops were reported to be able to ingest particles of a wide size range (particles up to 950 μm have been found in scallop stomachs; Mikulich and Tsikhon-Lukamina, 1981;

Shumway et al., 1987). Nevertheless, the assimilated Cd in the digestive gland was efficiently retained and was mainly bound to cytosolic compounds in the same proportions as in the food experiment, supporting the hypothesis of ingestion of sediment particles.

In the case of dietary exposure, Cd was assimilated to a similar extent in both species, with approx. 80% of the radiotracer being incorporated in the scallop tissues. Such a high assimilation efficiency (AE) is striking as in other bivalve species, lower values were generally reported, e.g. for the tropical clam *Gafrarium tumidum* (AE=42%), the tropical oysters *Isognomon isognomon* and *Malleus regula* (AEs=58 and 51%, respectively) and the blue mussel *Mytilus edulis* (AE ranging from 8 to 40%) (e.g., Wang and Fisher, 1997; Hédouin, 2006). These results suggest that food would be an important source of Cd for Pectinidae. However, inter-specific differences in Cd concentrations in scallops from the field (where *C. varia* showed the lowest concentrations) are difficult to explain in regards to the results obtained in our experiments. Indeed, lower depuration rates resulted in calculated biological half-life exceeding 3 years (Table 1), meaning that virtually all the assimilated Cd was readily stored in *C. varia* tissues. In contrast, the biological half-life following food exposure was approx. 4 months for *P. maximus*, indicating a faster turnover of the metal compared to *C. varia*. It is therefore likely that although living in the same areas, *C. varia* and *P. maximus* do not share the same food in the marine environment. Indeed, different storage mechanisms in prey can determine Cd bioavailability to higher trophic levels (e.g., Wallace and Lopez, 1997; Wallace and Luoma, 2003). Moreover, the dissolved and sediment pathways should also have a strong importance in *P. maximus* (see above). The use of a bioaccumulation model is therefore a mandatory step to further explore the importance of each exposure pathways (Thomann et al., 1995; Wang and Fisher, 1999). When applying such a model, food appears to be the major route of Cd accumulation in *C. varia*, with 99.6% of the metal being accumulated from phytoplankton. In *P. maximus*, it was not possible to determine accurate data for the model because the kinetic parameters of the post sediment-exposure loss phase were not significant. Therefore, we only considered food and seawater pathways. In such conditions, results indicated that food accounted for 84% of the accumulated Cd in *P. maximus*. Owing to the high assimilation efficiency of sediment-bound Cd (A_{01}=92%), it appears necessary to better delineate the sediment contribution to Cd accumulation in order to consider the three different pathways (seawater, food and sediment) on the global Cd bioaccumulation by *P. maximus*.

5. Conclusion

The present work on the bioaccumulation of Cd in two Pectinidae has confirmed the high Cd bioaccumulation potential of *C. varia* and *P. maximus*. The organs accumulating Cd to the highest extent in both species are the digestive gland and the kidneys whatever the exposure pathway was. Comparison of results from laboratory experiments clearly showed that *C. varia* showed higher bioconcentration and bioaccumulation capacities than *P. maximus*. Since field data have reported higher Cd levels in *P. maximus* than in *C. varia*, it is suggested that Cd should be bioaccumulated from sediment. Indeed, the high assimilation efficiency of Cd ingested through sediment pathway in *P. maximus* indicated that the particulate pathway could play an important role in the global Cd bioaccumulation process and studies on sediment as well as on suspended particulate matter should be further investigated to better simulate the different exposure routes of Cd to which Pectinidae are exposed in the field. Nevertheless, differences between field and laboratory observations could be related to different detoxification mechanisms in the two species.

Acknowledgements

The IAEA is grateful for the support provided to its Marine Environment Laboratories by the Government of the Principality of Monaco. This work was supported by the IAEA, the GIP Seine-Aval, the Conseil Général de la Charente-Maritime and the CRELA (Université de la Rochelle). We thank IFREMER and the Aquarium of La Rochelle for providing the scallops. We are grateful to L. Hédouin for her help in the modelling approach. MW is an Honorary Senior Research Associate of the National Fund for Scientific Research (NFSR, Belgium). **[SS]**

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