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## Spatial variability in total and organic mercury levels in Antarctic krill *Euphausia superba* across the Scotia Sea<sup>☆</sup>



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## ABSTRACT

Total and organic mercury concentrations were determined for males, females and juveniles of *Euphausia superba* collected at three discrete locations in the Scotia Sea (South Orkney Islands, South Georgia and Antarctic Polar Front) to assess spatial mercury variability in Antarctic krill. There was clear geographic differentiation in mercury concentrations, with specimens from the South Orkney Islands having total mercury concentrations 5 to 7 times higher than Antarctic krill from South Georgia and the Antarctic Polar Front. Mercury did not appear to accumulate with life-stage since juveniles had higher concentrations of total mercury ( $0.071 \mu\text{g g}^{-1}$  from South Orkney Islands;  $0.014 \mu\text{g g}^{-1}$  from South Georgia) than adults ( $0.054 \mu\text{g g}^{-1}$  in females and  $0.048 \mu\text{g g}^{-1}$  in males from South Orkney Islands;  $0.006 \mu\text{g g}^{-1}$  in females and  $0.007 \mu\text{g g}^{-1}$  in males from South Georgia). Results suggest that females may use egg laying as a mechanism to excrete mercury, with eggs having higher concentrations than the corresponding somatic tissue. Organic mercury makes up a minor percentage of total mercury (15–37%) with the percentage being greater in adults than in juveniles. When compared to euphausiids from other parts of the world, the concentration of mercury in Antarctic krill is within the same range, or higher, highlighting the global distribution of this contaminant. Given the high potential for biomagnification of mercury through food webs, concentrations in Antarctic krill may have deleterious effects on long-lived Antarctic krill predators.

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## 1. Introduction

Mercury contamination in the environment has been acknowledged as a global problem, the production and use of this element is nowadays very strictly regulated and limited (Selin, 2009; UNEP, 2013). Pathways of dispersion through ecosystems, including in the Antarctic, of this long-range contaminant are complex (Streets et al., 2009). Interplay between the distinctive Antarctic

atmosphere and the seasonal sea-ice cycle in the Southern Ocean generates a unique combination of environmental factors that can explain why the remote Southern Ocean has some of the highest reported concentrations of organic mercury (i.e. compounds containing covalent bonds between carbon and mercury) in open waters (Cossa et al., 2011). Due to its high affinity for proteins (Bustamante et al., 2006), organic mercury is the most toxic form of the element (Clarkson, 1992). It accumulates in aquatic organisms and biomagnifies within food webs, being toxic for top predators (Ackerman et al., 2014; Chouvelon et al., 2012; Coelho et al., 2010; Dehn et al., 2006) with consequences at the population level (Goutte et al., 2014a, 2014b). Wandering albatrosses, *Diomedea exulans*, are an example of this biomagnification effect in

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Antarctica, as it was found that they had some of the highest concentration of total mercury (from now on noted as mercury) in marine birds (up to  $24.80 \pm 8.61 \mu\text{g g}^{-1}$  dry weight) (Cherel et al., 2018; Tavares et al., 2013).

In the Southern Ocean, Antarctic krill, *Euphausia superba*, is a key species in the marine food webs connecting primary producers and higher predators (Everson, 2000). It has an estimated biomass of around 379 million tonnes (Atkinson et al., 2009) and being the main food for many vertebrates (Murphy et al., 2007; Xavier and Peck, 2015). For example, minke whales, *Balaenoptera acutorostrata*, and Crabeater seals, *Lobodon carcinophaga*, feed almost exclusively (>95%) on Antarctic krill (Adam, 2005; Armstrong and Siegfried, 1991; Croll and Tershy, 1998; Perrin et al., 2008). Chinstrap penguins, *Pygoscelis antarctica*, Gentoo penguins, *Pygoscelis papua*, and other species of penguins, in the Southern Ocean, also feed mostly on Antarctic krill (Dimitrijević et al., 2018; Xavier et al., 2018) with values around  $1.2 \text{ kg d}^{-1}$  (Croll and Tershy, 1998). Finally, Antarctic krill is the most harvested species in the Southern Ocean, with >260 000 tonnes fished in 2016, regulated under the Convention for the Conservation of Antarctic Living Resources (Nicol et al., 2000; CCAMLR, 2017).

In the context of environmental change (Constable et al., 2014; Cossa, 2013; Gutt et al., 2015), it is important to evaluate the impact of contaminants like mercury, particularly in a remote and presumably less impacted environments such as Antarctica with the associated risk to Southern Ocean top predators. This approach will contribute to a more in-depth knowledge of mercury bioaccumulation dynamics, in an effort towards the preservation of Antarctica ecosystems into the future (Rintoul et al., 2018; Seewagen, 2010). Despite the major role of Antarctic krill in the Southern Ocean, there are only a few studies reporting mercury concentrations in this region (Bargagli et al., 1998; Brasso et al., 2012b; Locarnini and Presley, 1995; Moren et al., 2006). Indeed, to our knowledge, no studies have ever analysed organic mercury content in Antarctic krill. Assessing the levels of organic mercury in such an important prey as Antarctic krill is crucial to better understand the pathway of this contaminant through Southern Ocean food webs. In this context, this study compares the total and organic mercury of Antarctic krill from three different locations: the South Orkney Islands, an Antarctic island group which experiences winter sea ice (Murphy et al., 1995); South Georgia, a sub-Antarctic island free of sea ice (Rogers et al., 2015); and the Antarctic Polar Front, a transition area from the Southern Ocean to the Atlantic Ocean with warmer waters (Dong et al., 2006). Under this context, differences among life stages (eggs, juveniles, adults) and sexes (males and females), were assessed and interpreted in the scope of a possible biomagnification of mercury in the Antarctic trophic web.

## 2. Material and methods

### 2.1. Sampling

Antarctic krill *Euphausia superba* were collected from the British research vessel RRS *James Clark Ross* during the austral summers of 2007/08, 2015/16 and 2016/17 (cruises JR177, JR15004 and JR16003 respectively). The three cruises sampled three areas of the Scotia Sea (Fig. 1) with different oceanic characteristics. JR16003 had one sampling point at the Antarctic Polar Front. Both JR16003 and JR177 sampled predominantly around South Georgia, and JR15004 sampled around the South Orkney Islands.

Samples were collected from the water column using an 8 m<sup>2</sup> mouth-opening Rectangular Midwater Trawl (RMT8; mesh size reducing from 4.5 mm to 2.5 mm in the cod end) (Roe and Shale, 1979). The net was rigged with two nets that could be remotely

opened and closed at different depths. The RMT8 was used to target particularly Antarctic krill swarms and other layers of interest (e.g. fish layers) identified by the vessel scientific echosounder system (i.e. Simrad EK60/EK80 operating between 38 and 200 kHz).

Antarctic krill in the catches were identified and total length (TL) of each individual was measured, from the anterior edge of the eye to the tip of the telson and rounded down (Morris et al., 1992). Sex and maturity stage were determined with reference to the presence of a petasma (males), thelycum (females) or absent (juveniles; individuals without visible external sexual characteristics) (Ross and Quetin, 2000). Samples were either preserved in sample bags at  $-20^\circ\text{C}$  (JR15004 and JR16003) or on vials in ethanol (for JR177) (Fort et al., 2016).

### 2.2. Laboratory procedures

Prior to the mercury analysis, all samples were freeze-dried for at least 24 h. The eggs of females (Maturity stage III) (Ross and Quetin, 2000) from JR177 (South Georgia) were removed under the microscope before freeze-drying.

Dried individuals and tissues were homogenized and analysed for total mercury by thermal decomposition atomic absorption spectrometry with gold amalgamation, using a LECO AMA-254 (Advanced mercury analyser) following Coelho et al. (2018). Organic mercury was determined through digestion with a mixture of 18% potassium bromide (KBr) in 5% sulfuric acid ( $\text{H}_2\text{SO}_4$ ), followed by extraction of organic mercury into toluene as described in Válega et al. (2006). Analytical quality control was performed using certified reference material (CRM; in this case TORT-2 and TORT-3 [lobster hepatopancreas, National Research Council, Canada]). The obtained values (mean  $\pm$  SD) for the whole of the CRM analyses ranged from 81 to 100% (TORT-2:  $87 \pm 3\%$ ,  $n = 41$ ; TORT-3:  $90 \pm 8\%$ ,  $n = 27$ ), results were corrected using the daily recovery efficiency of CRMs. The mass of CRM used for quality control analyses was adjusted to be within the range of total mercury (in ng) present in the samples. Analyses were performed in duplicate, blanks were analysed at the beginning of each set of samples and the coefficient of variation between replicates never exceeded 10%. CRMs were also used to validate organic mercury analyses, with an extraction efficiency of  $80 \pm 2\%$  and  $98 \pm 5\%$ , respectively. The limit of detection for this analytical method is  $0.01 \text{ ng g}^{-1}$  of total mercury and  $4 \text{ ng g}^{-1}$  for organic mercury. All concentration data are expressed subsequently in  $\mu\text{g g}^{-1}$  dry weight.

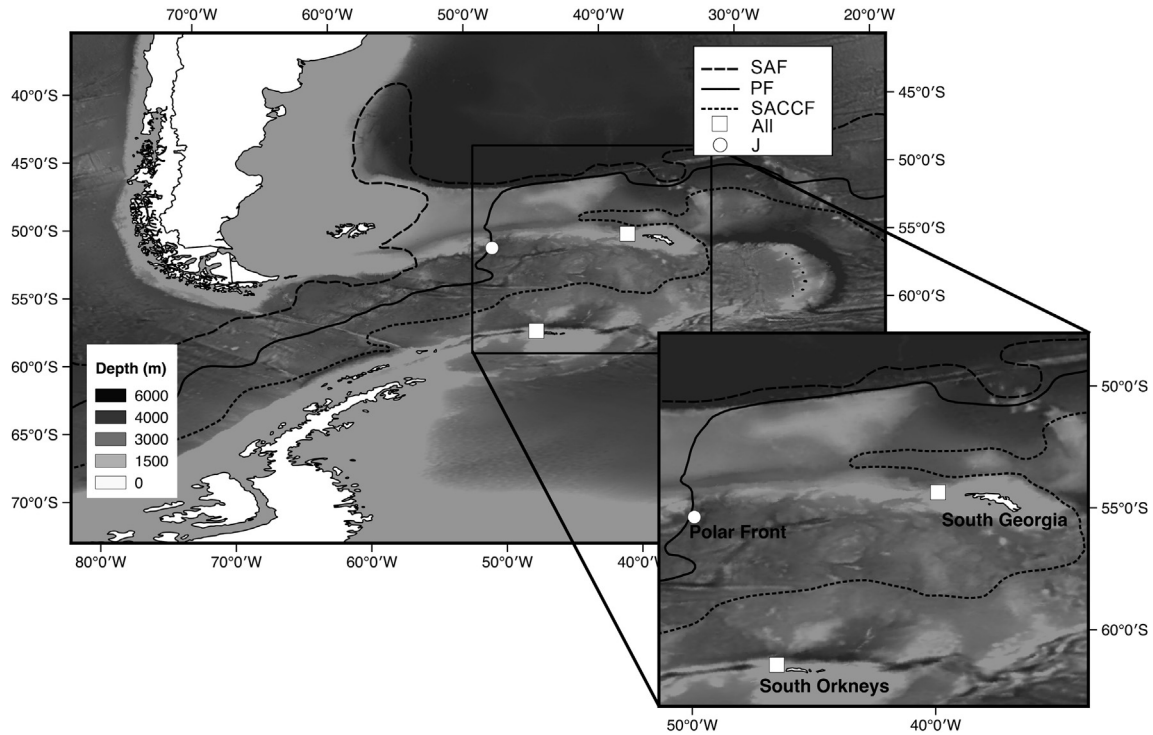
### 2.3. Statistical analysis

Wilcoxon tests were used to investigate whether there were any differences in mercury concentrations between females and males, between eggs and females, or between sampling sites. Kruskal-Wallis tests were performed to examine if there were statistical differences between sex/maturity and location. Linear regressions were calculated to examine possible relationships between Antarctic krill length and individual mercury concentration. All analyses were performed using the R software version 3.4.2 (R Core Team, 2013). All values are presented as mean  $\pm$  SD.

## 3. Results

### 3.1. Total mercury concentrations in Antarctic krill according to geographic areas

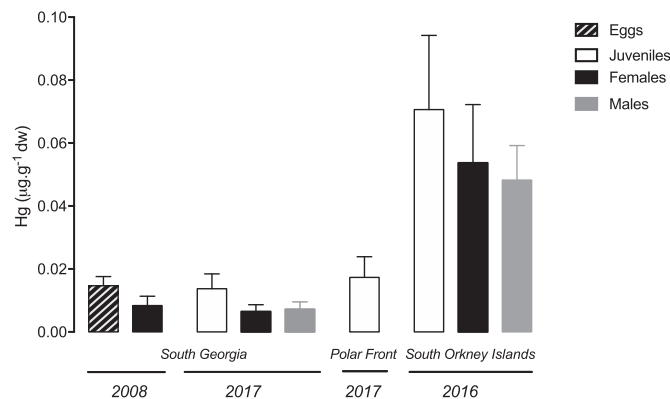
Total mercury concentrations varied between  $0.054 \pm 0.018 \mu\text{g g}^{-1}$  in females,  $0.048 \pm 0.011 \mu\text{g g}^{-1}$  in males and  $0.071 \pm 0.023 \mu\text{g g}^{-1}$  in juveniles from the South Orkney Islands to  $0.006 \pm 0.002 \mu\text{g g}^{-1}$  in females,  $0.007 \pm 0.002 \mu\text{g g}^{-1}$  in males and



**Fig. 1.** Sampling sites of Antarctic krill (white square "All" – samples of juveniles, females and males; white dot "J" – samples of juveniles) and general positions of the Subantarctic Front (SAF), Polar Front (PF) and the Southern boundary of the Antarctic Circumpolar Current Front (SACCF) (Sallé et al., 2008).

$0.014 \pm 0.005 \mu\text{g g}^{-1}$  in juveniles from the South Georgia and  $0.017 \pm 0.006 \mu\text{g g}^{-1}$  in juveniles from the Antarctic Polar Front.

There was a clear differentiation in mercury concentrations between the three locations (Fig. 2): adult Antarctic krill from the South Orkney Islands had concentrations of mercury about 7 times higher in females (Wilcoxon rank sum test,  $W = 120$ ,  $p < 0.001$ ) and males (Wilcoxon rank sum test,  $W = 120$ ,  $p < 0.001$ ) than adult Antarctic krill from South Georgia, and juveniles showed concentrations around 5 times higher in the South Orkney Islands (Kruskal-Wallis,  $H_3 = 41.03$ ,  $p < 0.001$ ) than those collected at South Georgia and at the Antarctic Polar Front. Juveniles from the northern locations (South Georgia and Antarctic Polar Front) had similar mercury concentrations (Wilcoxon rank sum test,  $W = 192$ ,  $p = 0.093$ ).



**Fig. 2.** Total mercury concentrations ( $\mu\text{g g}^{-1} \text{dw}$ ) in Antarctic Krill (*Euphausia superba*) collected around South Georgia in the austral summer of 2007/08 and 2016/2017, at the Antarctic Polar Front in the austral summer of 2016/17, and around the South Orkney Islands during the austral summer of 2015/16. Bars show the mean. Error bar is 1 standard deviation.

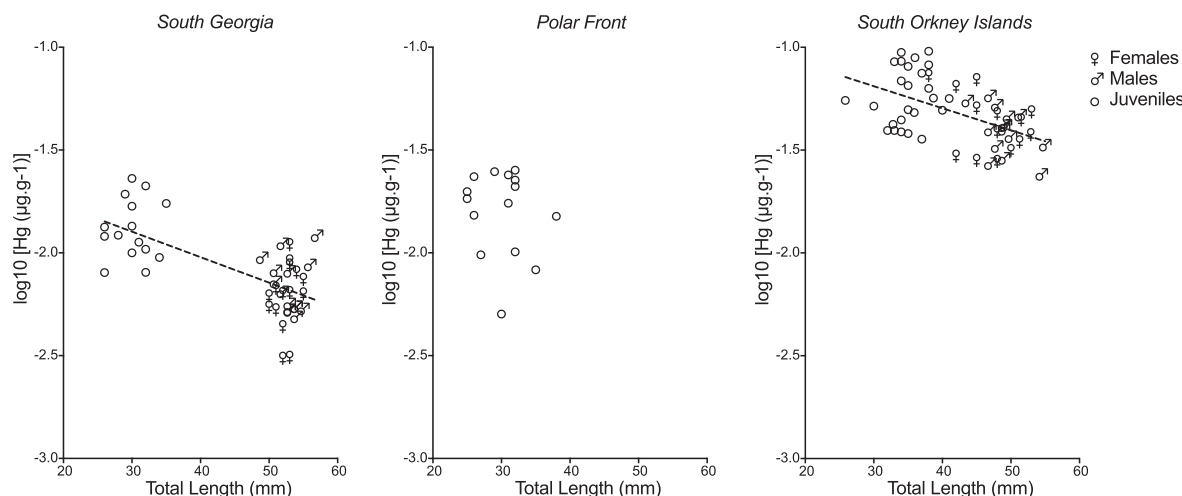
### 3.2. Total mercury concentrations in Antarctic krill according to life stage

There were significant differences (Wilcoxon signed rank test,  $Z = -3.351$ ,  $p = 0.001$ ) between the mercury concentrations in the eggs ( $0.015 \pm 0.002 \mu\text{g g}^{-1}$ ) and the corresponding female somatic tissue ( $0.008 \pm 0.003 \mu\text{g g}^{-1}$ ) from South Georgia (Fig. 2). There were no significant differences (Wilcoxon rank sum test,  $W = 189$ ,  $p = 0.071$ ) between the females sampled in 2007/08 and 2016/17 at South Georgia ( $0.007 \pm 0.002 \mu\text{g g}^{-1}$ ). Juveniles caught around South Georgia ( $0.014 \pm 0.005 \mu\text{g g}^{-1}$ ) had significantly higher mean concentration of mercury than adults ( $0.007 \pm 0.002 \mu\text{g g}^{-1}$ ; Kruskal-Wallis  $H = 41.031$ ,  $p < 0.01$ ) from the same region. Juveniles and eggs from South Georgia also had similar concentrations (Wilcoxon rank sum test,  $W = 205$ ,  $p = 0.254$ ). Like in juveniles from South Georgia, juveniles caught at the South Orkney Islands ( $0.071 \pm 0.024 \mu\text{g g}^{-1}$ ) also had significantly higher mercury concentrations than adults ( $0.051 \pm 0.015 \mu\text{g g}^{-1}$ ; Kruskal-Wallis  $H = 10.048$ ,  $p = 0.07$ ).

Significant negative correlations of mercury concentration with body size was common to both the South Orkney Islands and South Georgia ( $Y = -0.0124 * X - 1.525$ ,  $R^2 = 0.46$ ,  $F_{1, 43} = 36.41$ ,  $p < 0.001$  from South Georgia;  $Y = -0.01072 * X - 0.8675$ ,  $R^2 = 0.2746$ ,  $F_{1, 52} = 19.69$ ,  $p < 0.001$  from South Orkney Islands) meaning that bigger individuals had lower mercury concentrations (Fig. 3). It was not possible to discern if such a relationship also existed at the Antarctic Polar Front, since only juveniles were found at this location.

### 3.3. Total mercury concentrations in Antarctic krill according to sex

Concentrations of mercury in adult females ( $0.054 \pm 0.018 \mu\text{g g}^{-1}$ ) and males ( $0.048 \pm 0.011 \mu\text{g g}^{-1}$ ) from South Georgia were similar (Wilcoxon signed rank test,  $W = 216$ ,  $p = 0.5$ ; Fig. 2). There were also no differences in mercury concentration between



**Fig. 3.** Total mercury concentration ( $\mu\text{g g}^{-1}$  dw) on a log 10 scale versus total length (mm) for individual Antarctic krill (*Euphausia superba*) by maturity stage and sex respectively. Data are shown separately for krill collected around South Georgia ( $Y = -0.0124 \cdot X - 1.525$ ,  $R^2 = 0.46$ ,  $F_{1,43} = 36.41$ ,  $p < 0.001$ ), the Antarctic Polar Front (both in the austral summer of 2016/17) and the South Orkney Islands ( $Y = -0.01072 \cdot X - 0.8675$ ,  $R^2 = 0.2746$ ,  $F_{1,52} = 19.69$ ,  $p < 0.001$ ; summer of 2015/16).

sexes in the samples collected from the South Orkney Islands (Wilcoxon signed rank test,  $W = 199$ ,  $p=0.6$ ; Fig. 2).

3.4. Organic mercury in Antarctic krill

Adult Antarctic krill from the South Orkney Islands had higher concentrations of organic mercury than adults from South Georgia (Table 1) (for both males and females), but concentrations in juveniles were similar between the two locations. While no significant differences between juveniles, males and females were observed in the South Orkney Islands, juveniles in South Georgia had higher organic mercury concentrations than adults.

Organic mercury percentages in Antarctic krill were lower in the South Orkney Islands (15% in juveniles, 16% in females and 21% in males) than at South Georgia (29% in juveniles, 37% in females and 36% in males) and the Antarctic Polar Front (35% in juveniles; Table 1).

4. Discussion

Despite some studies reporting mercury levels in Antarctic krill (Bargagli et al., 1998; Brasso et al., 2012b; Locarnini and Presley, 1995; Moren et al., 2006), there has remained a gap in knowledge regarding variability in mercury concentration by size, gender and location. Furthermore, to our knowledge this is the first study to determine organic mercury concentrations in Antarctic krill.

4.1. Total mercury concentrations according to geographic areas

We found Antarctic krill from South Orkney Islands had mercury

body burdens 5 to 7 times higher than those from South Georgia and from the Antarctic Polar Front. Habitat differences may explain the variations in contamination levels between these three areas in the Southern Ocean. The average sea surface temperature around the South Orkney Islands is lower than in South Georgia (Barnes et al., 2005; Clarke and Leakey, 1996) and at the Antarctic Polar Front. This temperature gradient leads to an important ecosystem difference, promoting the presence of more winter ice in the South Orkney Islands (Atkinson et al., 2001). Ice formation can act as a buffer for mercury and other elements (Lindberg et al., 2002). Furthermore, the ice may act as a trap for contaminants precipitating from the atmosphere (Beyer and Matthies, 2001; Cossa et al., 2011), which are released into the water column upon ice melting (Brierley and Thomas, 2002; Geisz et al., 2008; Mastromonaco et al., 2017). In the Arctic, for instance, higher concentrations of mercury were measured in seawater under sea-ice, when compared with ice-free regions (Hintelmann et al., 2007) and higher concentrations of mercury were found under ice during spring (Mastromonaco et al., 2017). Additionally, depletion events promote higher precipitation rates of atmospheric mercury in colder areas, mainly during springtime, when halogen radicals oxidize the mercury (Ebinghaus et al., 2002; Lindberg et al., 2002). Indeed, these depletion events have been reported along and between regions of Antarctic sea-ice (Dommergue et al., 2010). Thus, higher depletion rates, sea ice formation and its melting may explain why there were more contaminants available to Antarctic krill around the South Orkney Islands than around South Georgia. Comparing our data with previous records of mercury in Antarctic krill, we see that samples from the Ross Sea, an area with winter sea ice (Bargagli et al., 1998), had higher concentrations than South

**Table 1**  
Organic mercury (OHg) and total mercury (THg) concentrations in samples of Antarctic krill (*Euphausia superba*) collected from different locations in the Scotia Sea during the austral summers of 2015/16 and 2016/17. Average  $\pm$  Standard deviation.

Location	Year	Sex/Maturity	Number	OHg ( $\mu\text{g g}^{-1}$ dw)	THg ( $\mu\text{g g}^{-1}$ dw)	OHg (%)
South Orkney Islands	2016	Juvenile	20	$0.008 \pm 0.003$	$0.051 \pm 0.016$	15%
South Orkney Islands	2016	Female	20	$0.008 \pm 0.002$	$0.052 \pm 0.022$	16%
South Orkney Islands	2016	Male	20	$0.008 \pm 0.003$	$0.040 \pm 0.014$	21%
South Georgia	2017	Juvenile	20	$0.008 \pm 0.002$	$0.024 \pm 0.006$	29%
South Georgia	2017	Female	20	$0.002 \pm 0.0002$	$0.006 \pm 0.0003$	37%
South Georgia	2017	Male	20	$0.003 \pm 0.0001$	$0.007 \pm 0.0004$	36%
Antarctic Polar Front	2017	Juvenile	20	$0.005 \pm 0.001$	$0.014 \pm 0.005$	35%

**Table 2**Total mercury concentrations ( $\mu\text{g g}^{-1}$  dw) in different species of Antarctic krill around the world from published data and this study (mean  $\pm$  standard deviation).

Species	Hg ( $\mu\text{g g}^{-1}$ )	Location	Reference
<i>Euphausia frigida</i>	0.023 $\pm$ 0.002	Kerguelen Islands	Cipro et al. (2018)
<i>Euphausia pacifica</i> & <i>Thysanoessa spinifera</i>	0.030	Californian Current	Sydeman and Jarman (1998)
<i>Euphausia superba</i>	0.008 $\pm$ 0.002	Antarctic Peninsula	Brasso et al., 2012a
<i>Euphausia superba</i>	0.008	Krill food	Moren et al., 2006
<i>Euphausia superba</i>	0.018 $\pm$ 0.005	King George Island	Cipro et al. (2016)
<i>Euphausia superba</i>	0.013 to 0.049	Antarctic Peninsula	Locarnini and Presley (1995)
<i>Euphausia superba</i>	0.077 $\pm$ 0.026	Ross Sea	Bargagli et al. (1998)
<i>Euphausia superba</i> (Adult)	0.007 $\pm$ 0.002	South Georgia	This study
<i>Euphausia superba</i> (Adult)	0.051 $\pm$ 0.015	South Orkney Islands	This study
<i>Euphausia superba</i> (Juvenile)	0.014 $\pm$ 0.004	South Georgia	This study
<i>Euphausia superba</i> (Juvenile)	0.017 $\pm$ 0.006	Polar Front	This study
<i>Euphausia superba</i> (Juvenile)	0.071 $\pm$ 0.023	South Orkney Islands	This study
<i>Euphausia triacantha</i>	0.036 $\pm$ 0.006	Kerguelen Islands	Cipro et al. (2018)
<i>Euphausia vallentini</i> (Large 25–30 mm)	0.017 $\pm$ 0.001	Kerguelen Islands	Cipro et al. (2018)
<i>Euphausia vallentini</i> (Small 16–24 mm)	0.042 $\pm$ 0.003	Kerguelen Islands	Cipro et al. (2018)
Euphausiidae	0.023 $\pm$ 0.004	Hudson Bay (Canada)	Foster et al. (2012)
<i>Meganyctiphanes norvegica</i>	0.130 $\pm$ 0.004	Arctic	Ritterhoff and Zauke (1997)
<i>Meganyctiphanes norvegica</i>	0.172 $\pm$ 0.014	Bay of Biscay	Chouvelon et al. (2012)
<i>Meganyctiphanes norvegica</i>	0.250	South of Portugal	Leatherland et al. (1973)
<i>Meganyctiphanes norvegica</i>	0.490	Mediterranean	Minganti et al. (1996)
<i>Thysanoessa inermis</i>	0.120 $\pm$ 0.004	Arctic	Ritterhoff and Zauke (1997)
<i>Thysanoessa</i> sp.	0.067 $\pm$ 0.031	Kerguelen Islands	Cipro et al. (2018)

Georgia and the Antarctic Peninsula (Brasso et al., 2012a; Cipro et al., 2016; Locarnini and Presley, 1995), but similar to those at the South Orkney Islands (Table 2).

Other possible explanations for the higher mercury contamination in Antarctic krill from the South Orkney Islands could be the proximity to active volcanoes, which are well-known sources of mercury (Varekamp and Buseck, 1981; Zambardi et al., 2009). Several volcanoes have recently been reported in the Antarctic Peninsula (van Wyk de Vries et al., 2018), which is closer to the South Orkney Islands than to the other two sampling sites in the present study. Nevertheless, the uptake of mercury from such sources is likely to be variable given that previous studies measuring mercury concentrations in Antarctic krill from the Antarctic Peninsula measured levels that were lower than those specifically in the South Orkney Islands Antarctic krill population reported here (Brasso et al., 2012a; Locarnini and Presley, 1995; Moren et al., 2006) (Table 2). Mercury body burdens in Antarctic krill may also be related to food availability (Chen and Folt, 2005). Phytoplankton blooms, which are a main source of mercury to krill, are spatially and temporally variable in the Southern Ocean and have a large influence on Antarctic krill growth (Atkinson et al., 2006; Cuzin-Roudy, 2000). Accordingly, the dynamics and availability of food between locations will probably have a significant effect on the mercury bioavailability, intake and bioaccumulation in Antarctic krill.

In comparison with other krill species around the world (Table 2), there are examples where the concentration of mercury is lower, for instance, species from the Order Euphausiacea in the Hudson bay (Canada) (Foster et al., 2012) and *Euphausia pacifica* in the Californian Current (Sydeman and Jarman, 1998) than in some of our samples. Mercury concentrations in euphausiids from more industrialized European regions (Chouvelon et al., 2012; Leatherland et al., 1973; Minganti et al., 1996) and the Arctic (Ritterhoff and Zauke, 1997) are nevertheless considerably higher than in Antarctic krill (Table 2). Higher concentrations are also evident in euphausiid populations in the sub-Antarctic Kerguelen Islands (Cipro et al., 2018) which, like the Southern Ocean, is likely to result from remote atmospheric sources (Cossa et al., 2011).

#### 4.2. Total mercury concentrations according to life stage and sex

Mercury concentrations in Antarctic krill unexpectedly

decreased with age (see results). Since juveniles have a faster rate of growth compared to adults, one would otherwise expect burdens to be lower in juveniles through a growth dilution effect, as reported for *Daphnia pulex* (Karimi et al., 2007). Furthermore, juveniles have more frequent moulting cycles compared to adults (Buchholz, 1991), and excretion ratios will probably be more efficient at these early stages. Somatic growth of Antarctic krill is pre-programmed to slow once a certain age or maturity has been reached (Tarling et al., 2006), in order to divert considerable resources to reproductive tissue when reaching adulthood (Atkinson et al., 2006; Cuzin-Roudy, 2000). Adults also prey on higher trophic levels compared to juveniles (Atkinson et al., 2002) which should mean higher biomagnification potential, and therefore contrary to what was observed. The higher contaminant load of juveniles when compared with adults has, however, been reported in previous studies on Antarctic krill (Locarnini and Presley, 1995) as well as in the subantarctic krill *Euphausia vallentini* (Cipro et al., 2018). One mechanism that may explain this phenomenon is through egg laying, which has been reported as an important elimination route for mercury in several organisms such as birds (Brasso et al., 2012a; Pedro et al., 2015) and fish (Johnston et al., 2001; Schofield et al., 1994), and also previously hypothesized for crustaceans species (Coelho et al., 2008). In the present study, the higher mercury concentrations were found in Antarctic krill eggs when compared to corresponding somatic tissue, suggesting that egg laying maybe an elimination mechanism. However, males also have lower mercury burdens compared to juveniles which either rules out this hypothesis or indicates that males also eliminate mercury through their own gonadic tissue. Spermatophores are regularly produced and passed out of the body throughout the lifespan of males, although concentrations of mercury in these structures has yet to be measured.

#### 4.3. Organic mercury

We found concentrations of the highly toxic, organic forms of mercury of between 0.002 and 0.018  $\mu\text{g g}^{-1}$  dw, with the higher concentrations being found in both the South Orkney Islands and South Georgia, particularly in juveniles. Antarctic krill is the main prey for several Southern Ocean predators and it is estimated that more than half of its total biomass of 379 Mt is eaten by whales,

seals, seabirds, squid and fish (Atkinson et al., 2009). Assuming the lowest individual mercury concentrations measured by the present study, this would mean 1.33 t of mercury will be passed on from the consumption of Antarctic krill, of which 0.57 t will be in the organic form. However, the 1.33t of mercury potentially transferred in the trophic web is a conservative number, as it was calculated from the lowest concentration levels found in the present study, that is, at the same time the lowest concentration ever measured in the literature. So, it can be considered an underestimation. This organic mercury will be potentially bioaccumulated in the tissues of Antarctic krill predators and transferred towards upper food web predators leading to its biomagnification. Thus, it may reach concentrations that can affect the behaviour, reproductive success and even to reduce the survival of the top predators (Tan et al., 2009; Eagles-Smith et al., 2018). Such bioaccumulation of organic mercury from Antarctic krill consumption can explain how some Antarctic seabirds have particularly high concentrations of mercury (Tavares et al., 2013).

## 5. Conclusions

The accumulation of mercury in Antarctic krill decreases with increasing body size and maturity. Juveniles have higher concentrations than adults which may be the result of mercury elimination through gonadic tissue (eggs and spermatophores).

The observed spatial differences suggest that Antarctic krill reflects differential contaminant bioavailability in the Southern Ocean, while further studies are needed to discern the most significant variables governing site-specific mercury bioaccumulation.

The range of mercury concentrations reported in Antarctic krill are within the same range, or even higher, than other euphausiids from areas closer to the industrialized part of the world, highlighting mercury as a global pollutant.

Overall, our results stress the need to put into action pollutant monitoring programs to evaluate the sources, pathways and effects of contaminants in remote ecosystems.

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## References

Ackerman, J.T., Eagles-Smith, C.A., Heinz, G., La Cruz, De, S.E.W., Takekawa, J.Y., Miles, A.K., Adelsbach, T.L., Herzog, M.P., Bluso-Demers, J.D., Demers, S.A., Herring, G., Hoffman, D.J., Hartman, C.A., Willacker, J.J., Suchanek, T.H., Schwarzbach, S.E., Maurer, T.C., 2014. Mercury in Birds of San Francisco Bay-Delta, California: Trophic Pathways, Bioaccumulation, and Ecotoxicological Risk to Avian Reproduction. Oakland, CA. <https://doi.org/10.3133/ofr20141251>.

Adam, P.J., 2005. Lobodon carcinophaga. Mamm. Species 772, 1–14. [https://doi.org/10.1644/1545-1410\(2005\)772\[0001:LC\]2.0.CO;2](https://doi.org/10.1644/1545-1410(2005)772[0001:LC]2.0.CO;2).

Armstrong, A.J., Siegfried, W.R., 1991. Consumption of antarctic krill by minke whales. *Antarct. Sci.* 3, 13–18. <https://doi.org/10.1017/S0954102091000044>.

Atkinson, A., Meyer, B., Stuübing, D., Hagen, W., Schmidt, K., Bathmann, U.V., 2002. Feeding and energy budgets of Antarctic krill *Euphausia superba* at the onset of winter—II. Juveniles and adults. *Limnol. Oceanogr.* 47, 953–966. <https://doi.org/10.4319/lo.2002.47.4.0953>.

Atkinson, A., Rothery, P., Tarling, G.A., 2006. Natural growth rates in Antarctic krill (*Euphausia superba*): II. Predictive models based on food, temperature, body length, sex, and maturity stage. *Limnol. Oceanogr.* 51, 973–987. <https://doi.org/10.4319/lo.2006.51.2.0973>.

Atkinson, A., Siegel, V., Pakhomov, E.A., Jessopp, M.J., Loeb, V., 2009. A re-appraisal of the total biomass and annual production of Antarctic krill. *Deep-Sea Res. Pt I* 56, 727–740. <https://doi.org/10.1016/j.dsr.2008.12.007>.

Atkinson, A., Whitehouse, M.J., Priddle, J., Cripps, G.C., Ward, P., Brandon, M.A., 2001. South Georgia, Antarctica: a productive, cold water, pelagic ecosystem†. *Mar. Ecol. Prog. Ser.* 216, 279–308. <https://doi.org/10.3354/meps216279>.

Bargagli, R., Monaci, F., Sanchez-Hernandez, J.C., 1998. Biomagnification of mercury in an Antarctic marine coastal food web. *Mar. Ecol. Prog. Ser.* 169, 65–76. <https://doi.org/10.3354/meps169065>.

Barnes, D.K.A., Linse, K., Waller, C., Morely, S., Enderlein, P., Fraser, K.P.P., Brown, M., 2005. Shallow benthic fauna communities of South Georgia Island. *Polar Biol.* 29, 223–228. <https://doi.org/10.1007/s00300-005-0042-0>.

Beyer, A., Matthies, M., 2001. Long-range transport potential of semivolatile organic chemicals in coupled air-water systems. *Environ. Sci. Pollut. Res. Int.* 8, 173–179. <https://doi.org/10.1065/esor2001.06.078>.

Brasso, R.L., Abel, S., Polito, M.J., 2012a. Pattern of mercury allocation into egg components is independent of dietary exposure in gentoo penguins. *Arch. Environ. Contam. Toxicol.* 62, 494–501. <https://doi.org/10.1007/s00244-011-9714-7>.

Brasso, R.L., Polito, M.J., Lynch, H.J., Naveen, R., Emslie, S.D., 2012b. Penguin eggshell membranes reflect homogeneity of mercury in the marine food web surrounding the Antarctic Peninsula. *Sci. Total Environ.* 439, 165–171. <https://doi.org/10.1016/j.scitotenv.2012.09.028>.

Brierley, A.S., Thomas, D.N., 2002. Ecology of Southern Ocean pack ice. *Adv. Mar. Biol.* 43, 171–276. [https://doi.org/10.1016/S0065-2881\(02\)43005-2](https://doi.org/10.1016/S0065-2881(02)43005-2).

Buchholz, F., 1991. Moulting cycle and growth of Antarctic krill *Euphausia superba* in the laboratory. *Mar. Ecol. Prog. Ser.* 69, 217–229. <https://doi.org/10.2307/24816834>.

Bustamante, P., Lahaye, V., Durnez, C., Churlaud, C., Caurant, F., 2006. Total and organic Hg concentrations in cephalopods from the North Eastern Atlantic waters: influence of geographical origin and feeding ecology. *Sci. Total Environ.* 368, 585–596. <https://doi.org/10.1016/j.scitotenv.2006.01.038>.

CCAMLR, 2017. Krill Fishery Report 2016. In: Convention on the Conservation of Antarctic Marine Living Resources.

Chen, C.Y., Folt, C.L., 2005. High plankton densities reduce mercury biomagnification. *Environ. Sci. Technol.* 39, 115–121. <https://doi.org/10.1021/es0403007>.

Cherel, Y., Barbraud, C., Lahournat, M., Jaeger, A., Jaquemet, S., Wanless, R.M., Phillips, R.A., Thompson, D.R., Bustamante, P., 2018. Accumulate or eliminate? Seasonal mercury dynamics in albatrosses, the most contaminated family of birds. *Environ. Pollut.* 241, 124–135. <https://doi.org/10.1016/j.envpol.2018.05.048>.

Chouvelon, T., Spitz, J., Caurant, F., Mèndez-Fernandez, P., Autier, J., Lassus-Débat, A., Chappuis, A., Bustamante, P., 2012. Enhanced bioaccumulation of mercury in deep-sea fauna from the Bay of Biscay (north-east Atlantic) in relation to trophic positions identified by analysis of carbon and nitrogen stable isotopes. *Deep-Sea Res. Pt I* 65, 113–124. <https://doi.org/10.1016/j.dsr.2012.02.010>.

Cipro, C.V.Z., Cherel, Y., Bocher, P., Caurant, F., Miramand, P., Bustamante, P., 2018. Trace elements in invertebrates and fish from Kerguelen waters, southern Indian Ocean. *Polar Biol.* 41, 175–191. <https://doi.org/10.1007/s00300-017-2180-6>.

Cipro, C.V.Z., Montone, R.C., Bustamante, P., 2016. Mercury in the ecosystem of admiralty bay, king George island, Antarctica: occurrence and trophic distribution. *MPB* 1–7. <https://doi.org/10.1016/j.marpolbul.2016.09.024>.

Clarke, A., Leahey, R.J.G., 1996. The seasonal cycle of phytoplankton, macronutrients, and the microbial community in a nearshore antarctic marine ecosystem. *Limnol. Oceanogr.* 41, 1281–1294. <https://doi.org/10.4319/lo.1996.41.6.1281>.

Clarkson, T.W., 1992. Mercury: major issues in environmental health. *Environ. Health Perspect.* 100, 31–38. <https://doi.org/10.1289/ehp.9310031>.

Coelho, J.P., Reis, A.T., Ventura, S., Pereira, M.E., Duarte, A.C., Pardal, M.A., 2008. Pattern and pathways for mercury lifespans bioaccumulation in *Carcinus maenas*. *Mar. Pollut. Bull.* 56, 1104–1110. <https://doi.org/10.1016/j.marpolbul.2008.03.020>.

Coelho, J.P., Santos, H., Reis, A.T., Falcão, J., Rodrigues, E.T., Pereira, M.E., Duarte, A.C., Pardal, M.A., 2010. Mercury bioaccumulation in the spotted dogfish (*Scyliorhinus canicula*) from the Atlantic Ocean. *Mar. Pollut. Bull.* 60, 1372–1375. <https://doi.org/10.1016/j.marpolbul.2010.05.008>.

Constable, A.J., Melbourne-Thomas, J., Corney, S.P., Arrigo, K.R., Barbraud, C., Barnes, D.K.A., Bindoff, N.L., Boyd, P.W., Brandt, A., Costa, D.P., Davidson, A.T., Ducklow, H.W., Emmerson, L., Fukuchi, M., Gutt, J., Hindell, M.A., Hofmann, E.E., Hsieh, G.W., Iida, T., Jacob, S., Johnston, N.M., Kawaguchi, S., Kokubun, N., Koubbi, P., Lea, M.-A., Makhado, A., Massom, R.A., Meiners, K., Meredith, M.P., Murphy, E.J., Nicol, S., Reid, K., Richerson, K., Riddle, M.J., Rintoul, S.R., Smith, W.O., Southwell, C., Stark, J.S., Sumner, M., Swadling, K.M., Takahashi, K.T., Trathan, P.N., Welsford, D.C., Weimerskirch, H., Westwood, K.J., Wienecke, B.C., Wolf-Gladrow, D., Wright, S.W., Xavier, J.C., Ziegler, P., 2014. Climate change and Southern Ocean ecosystems I: how changes in physical habitats directly affect marine biota. *Glob. Chang. Biol.* 20, 3004–3025. <https://doi.org/10.1111/gcb.12444>.

- doi.org/10.1111/gcb.12623.
- Cossa, D., 2013. Marine biogeochemistry: methylmercury manufacture. *Nat. Geosci.* 6, 810–811. <https://doi.org/10.1038/ngeo1967>.
- Cossa, D., Heimbürger, L.-E., Lannuzel, D., Rintoul, S.R., Butler, E.C.V., Bowie, A.R., Averty, B., Watson, R.J., Remenyi, T., 2011. Mercury in the Southern Ocean. *Geochim. Cosmochim. Acta* 75, 4037–4052. <https://doi.org/10.1016/j.gca.2011.05.001>.
- Croll, D.A., Tershy, B.R., 1998. Penguins, Fur seals, and fishing: prey requirements and potential competition in the South Shetland Islands, Antarctica. *Polar Biol.* 19, 365–374. <https://doi.org/10.1007/s003000050261>.
- Cuzin-Roudy, J., 2000. Seasonal reproduction, multiple spawning, and fecundity in northern krill, *Meganyctiphanes norvegica*, and Antarctic krill, *Euphausia superba*. *Can. J. Fish. Aquat. Sci.* 57, 6–15. <https://doi.org/10.1139/f00-165>.
- Dehn, L.-A., Follmann, E.H., Thomas, D.L., Sheffield, G.G., Rosa, C., Duffy, L.K., O'Hara, T.M., 2006. Trophic relationships in an Arctic food web and implications for trace metal transfer. *Sci. Total Environ.* 362, 103–123. <https://doi.org/10.1016/j.scitotenv.2005.11.012>.
- Dimitrijević, D., Paiva, V.H., Ramos, J.A., Seco, J., Ceia, F.R., Chipev, N., Valente, T., Barbosa, A., Xavier, J.C., 2018. Isotopic niches of sympatric Gentoo and Chinstrap Penguins: evidence of competition for Antarctic krill? *Polar Biol.* 1–15. <https://doi.org/10.1007/s00300-018-2306-5>.
- Dommergue, A., Sprovieri, F., Pirrone, N., Ebinghaus, R., Brooks, S., Courteaud, J., Ferrari, C.P., 2010. Overview of mercury measurements in the Antarctic troposphere. *Atmos. Chem. Phys.* 10, 3309–3319. <https://doi.org/10.5194/acp-10-3309-2010>.
- Dong, S., Sprintall, J., Gille, S.T., 2006. Location of the Antarctic polar front from AMSR-E satellite sea surface temperature measurements. *J. Phys. Oceanogr.* 36, 2075–2089. <https://doi.org/10.1175/JPO2973.1>.
- Eagles-Smith, C.A., Silbergeld, E.K., Basu, N., Bustamante, P., Diaz-Barriga, F., Hopkins, W.A., Kidd, K.A., Nyland, J.F., 2018. Modulators of mercury risk to wildlife and humans in the context of rapid global change. *Ambio* 47, 170–197. <https://doi.org/10.1007/s13280-017-1011-x>.
- Ebinghaus, R., Kock, H.H., Temme, C., Einax, J.W., Löwe, A.G., Richter, A., Burrows, J.P., Schroeder, W.H., 2002. Antarctic springtime depletion of atmospheric mercury. *Environ. Sci. Technol.* 36, 1238–1244. <https://doi.org/10.1021/es015710z>.
- Everson, I., 2000. *Krill: Biology, Ecology and Fisheries*. Blackwell Science, London.
- Fort, J., Grémillet, D., Traisnel, G., Amélineau, F., Bustamante, P., 2016. Does temporal variation of mercury levels in Arctic seabirds reflect changes in global environmental contamination, or a modification of Arctic marine food web functioning? *Environ. Pollut.* 211, 382–388. <https://doi.org/10.1016/j.envpol.2015.12.061>.
- Foster, K.L., Stern, G.A., Pazerniuk, M.A., Hickie, B., Walkusz, W., Wang, F., Macdonald, R.W., 2012. Mercury biomagnification in marine zooplankton food webs in Hudson bay. *Environ. Sci. Technol.* 46, 12952–12959. <https://doi.org/10.1021/es303434p>.
- Geisz, H.N., Dickhut, R.M., Cochran, M.A., Fraser, W.R., Ducklow, H.W., 2008. Melting glaciers: a probable source of DDT to the antarctic marine ecosystem. *Environ. Sci. Technol.* 42, 3958–3962. <https://doi.org/10.1021/es702919n>.
- Goutte, A., Barbraud, C., Meillère, A., Carravieri, A., Bustamante, P., Labadie, P., Budzinski, H., Delord, K., Cherel, Y., Weimerskirch, H., Chastel, O., 2014a. Demographic consequences of heavy metals and persistent organic pollutants in a vulnerable long-lived bird, the wandering albatross. *Proc. R. Soc. B* 281, 20133313. <https://doi.org/10.1098/rspb.2013.3313>.
- Goutte, A., Bustamante, P., Barbraud, C., Delord, K., Weimerskirch, H., Chastel, O., 2014b. Demographic responses to mercury exposure in two closely related Antarctic top predators. *Ecology* 95, 1075–1086. <https://doi.org/10.1890/13-1229.1>.
- Gutt, J., Bertler, N., Bracegirdle, T.J., Buschmann, A., Comiso, J., Hosie, G., Isla, E., Schloss, I.R., Smith, C.R., Tournadre, J., Xavier, J.C., 2015. The Southern Ocean ecosystem under multiple climate change stresses—an integrated circumpolar assessment. *Glob. Chang. Biol.* 21, 1434–1453. <https://doi.org/10.1111/gcb.12794>.
- Hintelmann, H., Graydon, J.A., Kirk, J.L., Barker, J., Dimock, B., Sharp, M.J., Lehnher, I., 2007. Methylated mercury species in Canadian high Arctic marine surface waters and snowpacks. *Environ. Sci. Technol.* 41, 6433–6441. <https://doi.org/10.1021/es070692s>.
- Johnston, T.A., Bodaly, R.A., Latif, M.A., Fudge, R.J., Strange, N.E., 2001. Intra- and interpopulation variability in maternal transfer of mercury to eggs of walleye (*Stizostedion vitreum*). *Aquat. Toxicol.* 52, 73–85. [https://doi.org/10.1016/S0166-445X\(00\)00129-6](https://doi.org/10.1016/S0166-445X(00)00129-6).
- Karimi, R., Chen, C.Y., Pickhardt, P.C., Fisher, N.S., Folt, C.L., 2007. Stoichiometric controls of mercury dilution by growth. *Proc. Natl. Acad. Sci. U.S.A.* 104, 7477–7482. <https://doi.org/10.1073/pnas.0611261104>.
- Leatherland, T.M., Burton, J.D., Culkin, F., 1973. Concentrations of some trace metals in pelagic organisms and of mercury in Northeast Atlantic Ocean water. *Deep-Sea Res.* 20, 679–685. [https://doi.org/10.1016/0011-7471\(73\)90085-5](https://doi.org/10.1016/0011-7471(73)90085-5).
- Lindberg, S.E., Brooks, S., Lin, C.J., Scott, K.J., Landis, M.S., Stevens, R.K., Goodsite, M., Richter, A., 2002. Dynamic oxidation of gaseous mercury in the Arctic troposphere at polar sunrise. *Environ. Sci. Technol.* 36, 1245–1256. <https://doi.org/10.1021/es0111941>.
- Locarnini, S.J.P., Presley, B.J., 1995. Trace element concentrations in antarctic krill, *Euphausia superba*. *Polar Biol.* 15, 283–288.
- Mastromonaco, M.G.N., Gärdfeldt, K., Assmann, K.M., Langer, S., Delali, T., Shlyapnikov, Y.M., Zivkovic, I., Horvat, M., 2017. Speciation of mercury in the waters of the Weddell, Amundsen and Ross seas (Southern Ocean). *Mar. Chem.* 193, 20–33. <https://doi.org/10.1016/j.marchem.2017.03.001>.
- Minganti, V., Capelli, R., De Pellegrini, R., Relini, L.O., Relini, G., 1996. Total and organic mercury concentrations in offshore crustaceans of the Ligurian Sea and their relations to the trophic levels. *Deep-Sea Res. I* 184, 149–162. [https://doi.org/10.1016/0048-9697\(96\)05076-0](https://doi.org/10.1016/0048-9697(96)05076-0).
- Moren, M., Suontama, J., Hemre, G.I., Karlsen, Ø., Olsen, R.E., Mundheim, H., Julshamn, K., 2006. Element concentrations in meals from krill and amphipods, — possible alternative protein sources in complete diets for farmed fish. *Aquaculture* 261, 174–181. <https://doi.org/10.1016/j.aquaculture.2006.06.022>.
- Morris, D.J., Ricketts, C., Watkins, J.L., Buchholz, F., Priddle, J., 1992. An assessment of the merits of length and weight measurements of Antarctic krill *Euphausia superba*. *Deep-Sea Res.* 39, 359–371. [https://doi.org/10.1016/0198-0149\(92\)90113-8](https://doi.org/10.1016/0198-0149(92)90113-8).
- Murphy, E.J., Clarke, A., Symon, C., Priddle, J., 1995. Temporal variation in Antarctic sea-ice: analysis of a long term fast-ice record from the South Orkney Islands. *Deep-Sea Res. I* 42, 1045–1062. [https://doi.org/10.1016/0967-0637\(95\)00057-D](https://doi.org/10.1016/0967-0637(95)00057-D).
- Murphy, E.J., Watkins, J.L., Trathan, P.N., Reid, K., Meredith, M.P., Thorpe, S.E., Johnston, N.M., Clarke, A., Tarling, G.A., Collins, M.A., Forcada, J., Shreeve, R.S., Atkinson, A., Korb, R., Whitehouse, M.J., Ward, P., Rodhouse, P.G., Enderlein, P., Hirst, A.G., Martin, A.R., Hill, S.L., Staniland, I.J., Pond, D.W., Briggs, D.R., Cunningham, N.J., Fleming, A.H., 2007. Spatial and temporal operation of the Scotia Sea ecosystem: a review of large-scale links in a krill centered food web. *Phil. Trans. R. Soc. B* 362, 113–148. <https://doi.org/10.1098/rstb.2006.1957>.
- Nicol, S., Forster, I., Spence, J., 2000. Products derived from krill. In: *Krill: Biology, Ecology and Fisheries*. Blackwell Science, London, pp. 262–283.
- Pedro, S., Xavier, J.C., Tavares, S., Trathan, P.N., Ratcliffe, N., Paiva, V.H., Medeiros, R., Vieira, R.P., Ceia, F.R., Pereira, E., Pardal, M.A., 2015. Mercury accumulation in gentoo penguins *Pygoscelis papua*: spatial, temporal and sexual intraspecific variations. *Polar Biol.* 1–11. <https://doi.org/10.1007/s00300-015-1697-9>.
- Perrin, W.F., Würsig, B., Thewissen, J., 2008. *Encyclopedia of Marine Mammals*. Academic Press, London. <https://doi.org/10.1016/B978-0-12-373553-9.X0001-6>.
- R Core Team, 2013. *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna.
- Rintoul, S.R., Chown, S.L., DeConto, R.M., England, M.H., Fricker, H.A., Masson-Delmotte, V., Naish, T.R., Siegert, M.J., Xavier, J.C., 2018. Choosing the future of Antarctica. *Nature* 558, 233–241. <https://doi.org/10.1038/s41586-018-0173-4>.
- Ritterhoff, J., Zauke, G.P., 1997. Trace metals in field samples of zooplankton from the Fram Strait and the Greenland Sea. *Sci. Total Environ.* 199, 255–270. [https://doi.org/10.1016/S0048-9697\(97\)05457-0](https://doi.org/10.1016/S0048-9697(97)05457-0).
- Roe, H.S.J., Shale, D.M., 1979. A new multiple rectangular midwater trawl (RMT 1+8M) and some modifications to the institute of oceanographic sciences' RMT 1+8. *Mar. Biol.* 50, 283–288. <https://doi.org/10.1007/BF00394210>.
- Rogers, A.D., Yesson, C., Gravestock, P., 2015. A biophysical and economic profile of South Georgia and the South Sandwich Islands as potential large-scale Antarctic protected areas. In: Curry, B. (Ed.), *Advances in Marine Biology*. Elsevier Ltd., pp. 1–286. <https://doi.org/10.1016/bs.amb.2015.06.001>.
- Ross, R., Quetin, L., 2000. Reproduction in Euphausiacea. In: *Krill*. Blackwell Science Ltd, Oxford, UK, pp. 150–181. <https://doi.org/10.1002/9780470999493.ch6>.
- Sallée, J.B., Speer, K., Morrow, R., 2008. Southern Ocean fronts and their variability to climate modes. *J. Clim.* 21, 3020–3039.
- Schofield, C.L., Munson, R., Holsapple, J., 1994. The mercury cycle and fish in the Adirondack lakes. *Environ. Sci. Technol.* 28, 136A–143A. <https://doi.org/10.1021/es00052a721>.
- Seewagen, C.L., 2010. Threats of environmental mercury to birds: knowledge gaps and priorities for future research. *Bird. Conserv. Int.* 20, 112–123. <https://doi.org/10.1017/S095927090999030X>.
- Selin, N.E., 2009. Global biogeochemical cycling of mercury: a review. *Annu. Rev. Environ. Resour.* 34, 43–63. <https://doi.org/10.1146/annurev.environ.051308.084314>.
- Streets, D.G., Zhang, Q., Wu, Y., 2009. Projections of global mercury emissions in 2050. *Environ. Sci. Technol.* 43, 2983–2988. <https://doi.org/10.1021/es802474j>.
- Sydesman, W.J., Jarman, W.M., 1998. Trace metals in seabirds, Steller sea lion, and forage fish and zooplankton from central California. *Deep-Sea Res. I* 36, 828–832. [https://doi.org/10.1016/S0025-326X\(98\)00076-9](https://doi.org/10.1016/S0025-326X(98)00076-9).
- Tan, S.W., Meiller, J.C., Mahaffey, K.R., 2009. The endocrine effects of mercury in humans and wildlife. *Crit. Rev. Toxicol.* 39, 228–269. <https://doi.org/10.1080/10408440802233259>.
- Tarling, G.A., Shreeve, R.S., Hirst, A.G., Atkinson, A., Pond, D.W., Murphy, E.J., Watkins, J.L., 2006. Natural growth rates in Antarctic krill (*Euphausia superba*): I. Improving methodology and predicting intermolt period. *Limnol. Oceanogr.* 51, 959–972. <https://doi.org/10.4319/lo.2006.51.2.0959>.
- Tavares, S., Xavier, J.C., Phillips, R.A., Pereira, M.E., Pardal, M.A., 2013. Influence of age, sex and breeding status on mercury accumulation patterns in the wandering albatross *Diomedea exulans*. *Environ. Pollut.* 181, 315–320. <https://doi.org/10.1016/j.envpol.2013.06.032>.
- UNEP, 2013. *Global Mercury Assessment 2013*. UNEP Chemicals Branch.
- van Wyk de Vries, M., Bingham, R.G., Hein, A.S., 2018. A new volcanic province: an inventory of subglacial volcanoes in West Antarctica. *Geol. Soc. London Spec. Publ.* 461, 231–248. <https://doi.org/10.1144/SP461.7>.
- Varekamp, J.C., Buseck, P.R., 1981. Mercury emissions from mount St Helens during September 1980. *Nature* 293, 555–556. <https://doi.org/10.1038/293555a0>.
- Válega, M., Abreu, S., Pato, P., Rocha, L., Gomes, A.R., Pereira, M.E., Duarte, A.C., 2006. Determination of organic mercury in biota, plants and contaminated sediments using a thermal atomic absorption spectrometry technique. *Water Air Soil Pollut.* 174, 223–234. <https://doi.org/10.1007/s11270-006-9100-7>.

- Xavier, J.C., Peck, L.S., 2015. Life beyond the ice. In: Exploring the Last Continent: an Introduction to Antarctica, Marine Ecosystems in the Southern Ocean. Springer International Publishing, Cham, pp. 229–252. [https://doi.org/10.1007/978-3-319-18947-5\\_12](https://doi.org/10.1007/978-3-319-18947-5_12).
- Xavier, J.C., Velez, N., Trathan, P.N., Cherel, Y., De Broyer, C., Cánovas, F., Seco, J., Ratcliffe, N., Tarling, G.A., 2018. Seasonal prey switching in non-breeding gentoo penguins related to a wintertime environmental anomaly around South Georgia. *Polar Biol.* 1–13. <https://doi.org/10.1007/s00300-018-2372-8>.
- Zambardi, T., Sonke, J.E., Toutain, J.P., Sortino, F., Shinohara, H., 2009. Mercury emissions and stable isotopic compositions at Vulcano Island (Italy). *Earth Planet. Sci. Lett.* 277, 236–243. <https://doi.org/10.1016/j.epsl.2008.10.023>.